Isolation, Identification and Characterization of Plastic Degrading Gut Bacteria from *Zophobas atratus* Larvae

By

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A thesis submitted to the Department of Mathematics and Natural Sciences in partial fulfillment of the requirements for the degree of Master of Science in Biotechnology

> Department of Mathematics and Natural Sciences Brac University April, 2024

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Declaration

It is hereby declared that

- 1. The thesis submitted is my original work while completing my degree at Brac University.
- 2. The thesis does not contain material previously published or written by a third party, except where this is appropriately cited through full and accurate referencing.
- 3. The thesis does not contain material that has been accepted, or submitted, for any other degree or diploma at a university or other institution.
- 4. I have acknowledged all main sources of help.

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Approval

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Ethics Statement

This research is done under proper supervision and it is the author's original work. No animal was harmed during experiments. The article is written in a manner that the write-up does not contain material previously published or written by a third party, except where this is appropriately cited through full and accurate referencing. The experiment was done by maintaining all the rules and regulations of the Biotechnology and Microbiology laboratory of the Department of Mathematics and Natural Sciences, BRAC University.

Abstract

Plastic pollution has become a major environmental concern globally, and novel and ecofriendly approaches like bioremediation are essential to mitigate the impact. This study investigated the biodegradation of three common plastic types, LDPE, LLDPE, and EPS, by *Zophobas atratus* larvae. Over 36 days, the average larval consumption was found to be 24.04% LDPE, 20.01% EPS and 15.12% LLDPE. FTIR analysis confirmed plastic oxidation in the gut. Gut bacteria were selectively isolated and identified as *Pseudomonas aeruginosa* strains. These bacteria showed the ability to degrade specific plastic types confirmed by SEM. Whole genome sequencing revealed many enzymes, along with virulence factors, antibioticresistance genes, and rhamnolipid biosurfactant biosynthesis genes in both isolates. Rhamnolipid analysis and AST were performed. This study indicated *Zophobas atratus* larvae as potential LDPE, LLDPE, and EPS biodegradation agents. Additionally, the isolated strains of *Pseudomonas aeruginosa* provide a more direct and eco-friendly solution for plastic degradation.

Keywords: *Zophobas atratus*; Linear low-density polyethylene; Low-density polyethylene; Expanded polystyrene; Biodegradation; *Pseudomonas aeruginosa*

Dedication

I dedicate this Thesis work to my family.

Acknowledgment

I am very much thankful to certain people for supporting me in this journey. First, I want to congratulate myself for completing the thesis successfully by abiding by the rules as much as possible. I am thankful to my parents, who have never stopped supporting me and were beside me in every moment of my life. I am very humbly indebted to my thesis supervisor, Professor Dr. M. Mahboob Hossain, Department of Mathematics and Natural Sciences, who guided me efficiently in this journey and without whom, the research would not be possible and helped me to find my true potential as a researcher. I would like to thank Professor Dr. A F M Yusuf Haider, Chairperson, MNS department for allowing me to do research on this topic. I am also thankful to Dr. Munima Haque, Associate Professor, MNS department for her helpful advices. I would like to thank Professor Dr. Aparna Islam, MNS department for motivating and supporting me throughout the journey. I would like to thank some of my lab colleagues, Shabnoor Binte Dayem, Research Assistant; Tamanna Islam Toma, Research Assistant; Nazifa Tabassum Tasnim, Research Assistant for having my back and supporting me. I would like to thank Rafeed Rahman Turjya, former lecturer, Brac University; Md Salman Shakil, lecturer, Brac University; Akash Ahmed; senior lecturer, Brac University for their support. I am also thankful to Mahruf Al Shahariar and Rezanur Rahman Howlader. I would like to also thank the lab assistants who made the lab work so much easier. I would like to show my gratitude to Dr. Muntasir Alam, Assistant Scientist, Emerging Infections, Infectious Diseases Division, icddr,b, and Afsana Rashed, Research Assistant, Infectious Disease Division, icddr,b for prompt response to my samples and sequencing.

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Chapter 1

Introduction

1.1 Bioremediation

Bioremediation, an environmentally friendly and advanced technique, utilizes natural biological processes to eliminate harmful contaminants from the environment, as defined by (Vidali, 2001). The escalating concern over environmental pollution worldwide is largely attributed to increased fossil fuel production and consumption (Omokhagbor Adams et al., 2020). In light of this pollution crisis, bioremediation emerges as a pivotal solution, leveraging microbes, fungi, insects, and green plants instead of chemical substances to restore environmental integrity. Bioremediation encompasses two main approaches: in-situ and exsitu. In-situ bioremediation treats pollutants directly at the site of contamination without excavation, offering cost-effective advantages over ex-situ methods. Addressing the urgent need for environmental sustainability requires embracing green technologies like bioremediation to remediate polluted ecosystems affected by human activities, industrialization, and agricultural practices (Arora, 2018; Juwarkar et al., 2010). Bioremediation, however, faces challenges, including the specific selection of microorganisms, the potential toxicity of biodegradation by-products, and slower kinetics compared to conventional methods (Abatenh et al., 2017). Despite these drawbacks, bioremediation holds promise as an effective and eco-friendly solution, necessitating further research to comprehend microbial degradation processes (Chatterjee et al., 2008). With its minimal environmental impact and suitability for deployment in contaminated areas, bioremediation stands as a vital technology in achieving sustainable development and mitigating global climate challenges.

1.2 Plastics

The term "plastic" is derived from the Greek word "plastikos," meaning "capable of being molded into various shapes." Plastics are comprised of carbon, hydrogen, silicon, oxygen, chloride, and nitrogen, and are primarily sourced from oil, coal, and natural gas. Polyethylene, a linear hydrocarbon polymer formed by long chains of ethylene monomers (C_2H_4) , constitutes approximately two-thirds, or precisely 64%, of all plastic (Goosey, 1985). Plastics play an indispensable role in the global economy, finding extensive application in agriculture, construction, healthcare, and various other sectors. They serve as the cornerstone of numerous industries due to their versatility in manufacturing a wide array of products, ranging from household items to defense components. Additionally, plastics are utilized in the packaging of cosmetics, detergents, and pharmaceuticals. However, the excessive use of plastics poses a significant threat to both the environment and human well-being. The accumulation of plastics on land and in oceans has spurred considerable interest in the degradation of these polymers. To mitigate the adverse environmental effects of plastics, biodegradation methods are imperative. Understanding the interaction between microbes and polymers is critical in addressing plastic-related challenges. Many organisms, predominantly bacteria, have evolved strategies for the survival and decomposition of plastics (Oliveira et al., 2020). This study was focused on three different types of plastics, including Low-Density Polyethylene (LDPE), Expanded Polystyrene (EPS), and linear Low-Density Polyethylene (LLDPE).

1.3 Bioremediation of Plastics

Non-biodegradable plastics persist for centuries, ultimately transforming into contaminating microplastics that reenter food chains (Chamas et al., 2020). Their presence has become widespread in our air, water, soil, and food items (Wright et al., 2021). In 2017, global plastic emissions amounted to 0.8 million tons (mt) of microplastics and 8.7 mt of macroplastics; by 2050, this emission may increase to 2.2 gigatonnes (Gt) and 3.1 Gt respectively (Schwarz et al., 2023). These findings indicate the severity of plastic pollution and necessitate the development of safe, rapid, and effective plastic remediation methods. With this in mind, there is a wide range of commercially available plastics with different utilities, such as polyethylene (PE) - which can be of high, low, linear density, expanded polystyrene (EPS), polyethylene terephthalate (PET), polyvinyl chloride (PVC), and polypropylene (PP), etc. (Landrock, 1995). Among them, linear low-density polyethylene (LLDPE), low-density polyethylene (LDPE), and expanded polystyrene (EPS) are three of the most used plastic types in the world (Chamas et al., 2020). PE is a type of polyolefin that has a chemical formula of $(C_2H_4)_{n}$, which is a polymer of ethylene (or ethene) monomer produced by addition or radical polymerization by Ziegler-Natta polymerization or metallocene catalysis methods (Landrock, 1995). In this study, two types of PE were used, LDPE and LLDPE. LDPE is a type of branched PE, which has a high degree of short-chain branching along with long-chain branching with a density of 0.917- 0.940 $g/cm³$. It has low crystallinity, is highly amorphous, and has very low water absorption capability (Landrock, 1995). On the other hand, LLDPE is a linear polymer with significant numbers of short branches with a density of 0.915-0.950 g/cm³. Both LDPE and LLDPE are used to manufacture grocery bags, garbage bags, packaging film, agricultural mulch, insulation for wires and cables, bottles, toys, housewares, etc. Similarly, EPS is a white foam plastic material produced from solid beads of polystyrene with the chemical formula (C_8H_8) _n and a

density of 0.012-0.05 g/cm³ (Landrock, 1995). It is used in foam packaging, CD and DVD cases, insulation, peanuts for shipping, food packaging, meat/poultry trays, and egg cartons.

Plastic pollution has become a global problem due to inadequate recycling compared to its widespread use. Currently, chemical methods and bioremediation are used for remediating environmental plastics. Unfortunately, chemical remediation methods have some negative effects on the environment. On the other hand, bioremediation processes use microorganisms like bacteria, fungi, algae or insects to degrade, remove, change, immobilize, or detoxify pollutants from the environment, which is an eco-friendly alternative (Omokhagbor Adams et al., 2020)**.** Characteristics of the targeted plastics like mobility, crystalline structure, molecular weight, functional groups and additives, etc. influence the effectiveness of bioremediation. To undergo the process, microorganisms need to adhere to the surface of the plastics, followed by colonization, conversion of polymers to monomers, and finally monomers to simple compounds like CO2, water, ethylene glycol, etc. (Shah et al., 2008). Enzymatic degradation is one of the main mechanisms of these conversions when hydrolytic enzymes such as cutinase, lipase, proteinase K, dehydrogenase, etc. perform hydrolysis (Mohee and Unmar, 2007)**.**

Multiple studies have aimed to find suitable microorganisms to degrade different types of plastics. For example, *Brevibacillus borstelensis* and *Rhodococuus ruber* were identified as potential LLDPE-degrading microorganisms (Hadad et al., 2005; (Orr) et al., 2004). Moreover, PS-degrading bacteria like *Acinetobacter* sp. (Wang et al., 2020), *Serratia marcescens*, *Pseudomonas* sp., *Bacillus* sp. (Galgali et al., 2002), and LDPE-degrading bacteria like *Pseudomonas* sp. (Rajandas et al., 2012; Tribedi and Sil, 2013), *Bacillus amyloliquefaciens* (Nowak et al., 2011) have been isolated. In the meantime, other enzymes have also been identified that facilitate plastic degradation, such as alkane monooxygenase, laccase, and alkane hydroxylase (Bardají et al., 2019; Kim et al., 2021a; Santo et al., 2013).

Insects can also be used for plastic degradation. Since there is limited necessity to pre-condition plastics, it is possible to save time and money. Insects can act on their own and adapt to the changes in the environment. Moreover, they break down the plastics during consumption, which helps bacteria to degrade plastic better. Sometimes the gut bacteria gets additional assistance in the degradation from the enzymes secreted by the insects. For this advantage, extensive research is ongoing in this field, as multiple insect species have shown the capacity to degrade plastic. Biodegradation of PS by the larvae of *Tenebrio molitor* Linnaeus (mealworms) was first reported in 2015 (Yang et al., 2015a, 2015b). The same larvae were also observed to perform LDPE biodegradation (Brandon et al., 2018). Another insect larvae, *Galleria mellonella*, also known as greater wax moths, has been shown to degrade LDPE (Bombelli et al., 2017) and PS (Lou et al., 2020). *Galleria mellonella* has two enzymes in their saliva - 'Demetra', an arylphorin, and 'Ceres', a hexamerin - which can degrade PE within a few hours at room temperature (Sanluis-Verdes et al., 2022). Furthermore, *Tenebrio obscurus* (dark mealworm) is also reported to degrade PS (Peng et al., 2019) and LDPE (Yang et al., 2021b). Lastly, *Zophobas atratus* (synonymous with *Z. morio*, Coleoptera: Tenebrionidae) is another well-known insect that can degrade several types of plastics. This insect, also known as "Superworm", is a type of darkling beetle. This species is known to be a good nutrient source for livestock animals and aquaculture (Jabir et al., 2012; Rumbos and Athanassiou, 2021). Previous studies have revealed that at the larval stage, this insect can biodegrade plastics of different types (Sun et al., 2022)**,** such as PS and LDPE (Peng et al., 2020)**,** and polypropylene (Yang et al., 2021a). Most of these studies were done with LDPE and PS, but not with LLDPE (Peng et al., 2020; Yang et al., 2020; Zielińska et al., 2021). Moreover, although there is available data on the plastic-consuming capability of this insect from a few countries, no such data is available for Bangladesh. Interestingly, studies strongly indicate that their gut microbiome is connected with the biodegradation of plastics (Sun et al., 2022; Yang et al.,

2020). Studies show that the gut microbiome associated with plastic degradation includes genera from *Pseudomonas*, *Rhodococcus,* and *Corynebacterium* (Sun et al., 2022). Genomic data is absent for the potential plastic-degrading bacteria that constitute this microbiome.

The focus of this study revolved around the degradation of LDPE, LLDPE, and EPS, as these plastic types are produced commercially and used indiscriminately in Bangladesh. The bioremediation capacity of *Z. atratus* provides a potentially sustainable solution for the widespread effect of these plastics on the environment. Additionally, this study focused predominantly on the *Pseudomonas* genus, as they are known for their ability to survive in xenobiotic environments and their plastic-degrading capabilities (Wilkes & Aristilde, 2017; Wasi *et al.*, 2013; Lee *et al.*, 2020). For the experiments, the plastic degradation capability of the locally cultivated larvae was assessed on LDPE, LLDPE, and EPS samples collected from markets. After feeding on the plastics, consumed plastic and larval frass were analyzed by Fourier Transform Infrared Spectroscopy (FTIR). Potential bacteria were isolated from the frass, and Scanning Electron Microscopy (SEM) was performed to prove the biodegradation of plastic feed. Whole Genome Sequencing was performed on the isolated bacteria followed by annotation. This study aimed to establish the plastic-degrading capability of the larvae at room conditions as well as gather and analyze the genomic data of the insects' gut bacteria which are involved in the degradation**.**

1.4 *Zophobas atratus*

Zophobas atratus, commonly known as the superworm, is renowned for its substantial size, feeding capacity, and both biological and economic potential. Previous studies have highlighted its nutritional value, making it a valuable nutrient and antimicrobial source for poultry feed. Recent research has also revealed its potential for waste management as a bioremediation agent (Rumbos & Athanassiou, 2021).

Taxonomic classification of *Z. atratus* has been subject to confusion, with recent research suggesting it to be conspecific with *Zophobas morio*, formerly known as *Tenebrio atratus*, and *Zophobas rugipes* (Tschinkel, 1984). *Z. atratus* belongs to the large beetle family Tenebrionidae (Park et al., 2013). Throughout its lifecycle, *Z. atratus* typically undergoes four distinct stages: Eggs, Larvae, Pupa, and Adults (Rumbos & Athanassiou, 2021). Female *Z. atratus* lay a significant number of oval-shaped eggs (approximately 2200) during their lifetime, each measuring around 1.7 mm in length and 0.7 mm in width (Fursov & Cherney, 2018). Larvae are typically cylindrical, about 55 mm long, with a sclerotized exoskeleton and 7 to 9 abdominal segments. The subsequent pupal stage lasts approximately 13-15 days, during which the pupae exhibit primarily quiescent behavior, although they can display defensive responses such as rotating the abdominal portion (Ichikawa & Kurauchi, 2009). Upon completing the pupal stage, *Z. atratus* emerges as an adult with a body length ranging from 38 to 57 mm, characterized by an elongated body and filiform antennae, with an average lifespan of about 180 days (Fursov & Cherney, 2018).

The larvae stage of *Z. atratus* holds significant biological and commercial importance, particularly as animal feed. In certain ethnic groups in Mexico, species of *Zophobas* are consumed as food (Ramos-Elorduy, 2009). In Brazil, *Z. atratus*is considered a potential protein and nutrient source for livestock and aquaculture feed (Soares Araújo et al., 2019). Recent studies have indicated that the plastic degradation capability of *Z. atratus* primarily resides in its gut microbiome, with certain bacterial strains, such as *Pseudomonas*, being isolated for further investigation toward utilizing *Z. atratus* as a potential bioremediation agent (Peng et al., 2020; Y. Yang et al., 2020). When the gut microbiome was treated with antibiotics, the plastic degradation ability of the superworm was significantly reduced, underscoring the contribution of the gut microbiome to this capability (Peng et al., 2020; Y. Yang et al., 2020).

1.5 Fourier Transform Infrared Spectroscopy (FTIR)

FTIR spectroscopy operates on the principle of interference between two beams of radiation, resulting in the generation of an interferogram. This signal is produced based on changes in path length between the beams reflected from mirrors within the interferometer block. Through Fourier transformation, distance and frequency domains are mathematically interconverted, giving rise to the name Fourier Transform Infrared Spectroscopy.

The key distinction between an FTIR spectrometer and a dispersive IR spectrometer lies in the use of the Michelson interferometer. The Michelson interferometer serves as the central component of FTIR spectrometers, dividing one light beam into two to create distinct paths. Subsequently, it combines the beams before directing them to the detector, where intensity differences between the two beams are measured relative to path variations. Essential components of FTIR include IR sources, detectors, beam splitters, and Fourier transforms. Third-generation FTIR spectrometers offer notable advantages, including significantly improved signal-to-noise ratios compared to earlier generations, rapid scanning of all frequencies (approximately 1 second), and exceptionally high resolution $(0.1 \sim 0.005 \text{ cm}^{-1})$. They also boast high accuracy in wave number measurements and a wide scan range $(1000 \sim$ 10 cm-1), along with reduced interference from extraneous light sources.

However, limitations of FTIR spectroscopy include the compact size of the sampling chamber, which restricts the size of samples that can be analyzed. Additionally, obstructive mounted pieces may interfere with the IR beam, further limiting sample size. Certain materials may completely absorb infrared radiation, rendering measurement impossible in such cases (*FTIR* *Micro-Spectrometer - BDD : Industrial Synchrotron Light Research Institute (Public Organization)*, n.d.)

1.6 Scanning Electron Microscopy (SEM)

The scanning electron microscope (SEM) utilizes electrons, rather than light, to generate images. Since their inception in the early 1950s, SEMs have significantly advanced research across various disciplines in the medical and physical sciences by enabling the examination of a broader range of specimens.

Compared to traditional microscopes, SEMs offer several advantages. They possess a large depth of field, allowing for more of a specimen to remain in focus simultaneously. Additionally, SEMs boast higher resolution capabilities, enabling the magnification of closely spaced specimens at greater levels. Unlike optical lenses, SEMs utilize electromagnets, granting researchers greater control over magnification levels. These advantages, coupled with the production of remarkably clear images, establish the SEM as a highly valuable research instrument.

Operationally, the SEM functions by producing a highly magnified image through the utilization of an electron beam emitted from an electron gun situated at the top of the microscope. This electron beam traverses a vertical path within a vacuum environment contained within the microscope. Along its path, the beam passes through electromagnetic fields and lenses, which focus it toward the specimen. Upon contact with the specimen, electrons and X-rays are emitted. Detectors within the SEM capture these emitted X-rays, backscattered electrons, and secondary electrons, converting them into a signal that is transmitted to a screen akin to a television screen, ultimately producing the final image (*Scanning Electron Microscope - Environmental Health and Safety - Purdue University*, n.d.).

Chapter 2

Materials and Methods

2.1 Working Place for the Study

The present research work was performed in the Biotechnology and Microbiology Laboratory of the Department of Mathematics and Natural Sciences, BRAC University, KHA 224, Progati Sarani, Merul Badda, Dhaka 1212.

2.2 Media, Solutions, and Reagents

Media, reagents, and solutions that were used in this thesis work were available as a reagent grade, and without further purification, those were used.

2.3 Handling of Laboratory Equipment

Detergents were used to wash all the glassware and rinsed 4-5 times with tap water. Autoclavable equipment was sterilized by autoclaving at 121° C for 15 minutes at 15 psi. All the microbiological works were done inside the Biological Safety Cabinet. Larvae were kept in PET boxes which were in a hardboard box and safety was maintained so that no larvae could leave the box.

2.4 Sample collection

About 300 pieces of *Zophobas atratus* larvae were bought from "Green Field Agro", a commercial cultivator from Pallabi, Mirpur, Dhaka, Bangladesh who was breeding the larvae for use as poultry/reptile feed, in a PET box with rice bran to feed them. The larvae were identified based on their morphology and darker color.

2.5 Preparation for Larvae Sample

2.5.1 Plastic Preparation

Three types of plastic were chosen, LDPE, LLDPE, and EPS. These plastics were cut into square or rectangular shapes, measured by weight, and put in three different empty PET boxes. The same process was done for the main process after 14 days.

For bacteriological studies, LDPE, LLDPE, and EPS were cut into 1 x 1 cm size and transferred into a beaker with distilled water and stirred for another 10 minutes. This step was repeated 3 times until all plastics were ridden of any residual surface dust. Then, they were aseptically placed in a 70% ethanol solution for 30 minutes. Finally, the disinfected plastics were transferred to a sterile petri dish dried in the laminar hood, and put away for further use.

2.5.2 Larvae Preparation

At the start of the study, the average weight of the larvae was 388 ± 23 mg. There was confirmation from the cultivators that the larvae had no antibiotics in their system as they were given antibiotic-free feeds. The absence of antibiotics was important, as their use in the feed would inhibit gut bacteria which were explored in this study. The organic feed used in this study was locally sourced rice bran.

2.6 Plastic Biodegradation Rate and Larvae Survivability Analysis

Z. atratus larvae were subjected to plastic consumption tests followed by survivability rate (SR) tests. The larvae were only fed rice bran for 18 days after collection from the cultivator. All the larvae were kept in the same environmental condition with the same rice bran provided as feed. After 18 days, three PET boxes were used to house the larvae, ensuring that no light could penetrate inside. Although PET itself is a type of plastic, *Z. atratus* larvae have not shown PET-degradation capability; hence PET boxes were used in this study. Each box was designated for a different plastic-type, and it was ensured that there was no edible material in the box that the larvae could feed on. The lids of the boxes had small holes for the circulation of air. For each plastic type, about 20 larvae were chosen randomly and put in the designated box. Then for the next 14 days, the larvae were kept with the plastic as the only food source to clear out the digestive system of any previous organic food. For positive control (PCN), a separate PET box was designated and 20 randomly selected larvae were kept inside with rice bran as a food source. Similarly, for negative control (NCN), another PET box was designated and 20 randomly selected larvae were kept inside with no food source. The frass was removed from the boxes every two days. After 14 days, all the used plastics and frass (rice bran for the PCN) were removed. The larval live average weight was measured for every box at this stage. Then, fresh weighted plastics were introduced according to the box designation (only rice bran for PCN) and data was collected over the next 36 days. The boxes were put in a bigger container which again was ensured to have no light penetration. The experiment was done at room temperature with no commercial incubator. For 36 days, the weight of the plastics was measured every 2 days to determine how much plastic was being eaten by the larvae. It was repeatedly ensured that no outside organic food source was available in the boxes. Additionally, frass was removed every 2 days and average larvae live weight was measured in the meantime. A live larvae count was carried out every day to check how many larvae have survived. Live larvae weight measurement was done every 2 days for the full length of the experiment (36 days) and onwards. However, after the first 15 days, frass was not removed for 7 days so that enough frass could be collected for bacteriological and chemical experiments. Then frass removal was again resumed every 2 days.

Plastic consumption data was calculated as follows:

Plastic Consumption (PC)
$$
(\%) = (P_{36}/P_0)*100\%
$$

Where P_{36} is the total consumed plastic after 36 days and P_0 is the initial weight of the plastic. Larvae survivability data was calculated as follows:

Survivalivity Rate (SR)
$$
(\%) = (S_{36}/S_0) * 100\%
$$

Where S_{36} is the total live larvae remaining after 36 days and S_0 is the initial larvae amount.

Lastly, for larvae average weight was calculated as follows:

Larvae Average Weight
$$
(LW) = W_x/N_x
$$

Where W is the combined weight of all the larvae in a box at x day, and N is the number of live larvae at x day. All the experiments were done in triplicate and the final result was formed with the average of the triplicate results.

2.7 Frass Bacteria Culture, Isolation, and Identification

By the 22nd day of the experiment, enough frass was accumulated in the PET boxes, so it was collected. Frass collected from the same designated boxes was mixed to prepare three master frass stocks (one each for the different plastic types). From each of the master stocks, 0.1g of frass was mixed in 10 ml 0.9% NaCl solution, and serial dilution was done up to 10^{-4} dilution. After dilution, 2.5 ml of each sample was inoculated in 250 ml of Minimal Salt (MS) broth. The MSB media composition was as follows: $KH_2PO_4(3g/L)$, $K_2HPO_4(0.1g/L)$, NaCl (5g/L), NH₄Cl (2g/L), MgSO₄.7H₂0 (0.16g/L), CaCl₂.2H₂0 (0.1g/L), with pH adjusted to 7.0 (Fazito do Vale et al., 2007). The media was sterilized by autoclaving at 121°C with 15 psi for 15 minutes. MS broth contained only salt and no carbon source, so plastics as a sole carbon source could be added to the media after inoculation to select only plastic-degrading bacteria.

LDPE, LLDPE, and EPS were cut and subjected to sterilization to ensure no contamination with carbon sources other than plastics according to the method described before. The weight of the plastics was measured before putting them into the designated sample. For positive control (PCB), MS medium supplemented with 0.1% glucose was used. For negative control (NCB), an MS medium with no carbon source was added. Inoculation for the PCB and NCB was done by mixing all the diluted samples $(3.3*3 \text{ ml})$ in 250 ml MS medium. The incubation

was done in a shaker incubator at 130 rpm at 37°C. The incubation period was 60 days, and bacterial growth was observed every 7 days.

After 60 days of incubation, $100 \mu L$ of the incubated sample was cultured on Nutrient Agar (NA), MacConkey agar, Cetrimide agar, and Mannitol Salt Agar (MSA) using the spread plate method. NA can support a wide range of non-fastidious bacteria and is used as an indicator for the presence of bacteria in the incubated MS broth. The other three media were used due to the focus on *Pseudomonas* - MacConkey agar and Cetrimide agar can selectively grow *Pseudomonas*, whereas MSA does not support its growth (Brown & Lowbury, 1965). The other purpose of MSA was to evaluate if any gram-positive bacteria were growing in MS broth. The bacterial colonies that grew were first isolated and differentiated via colony morphology and selective growth. Then, depending on the different morphology observed among the three designated plastic types, polymerase chain reaction (PCR) was done to confirm the *Pseudomonas* genus with PA-GS primer pairs. For the PCR process, conditions were as described by (Spilker et al., 2004).

In the PCR process, the primer pair used was PA-GS-F and PA-GS-R (Table 1). A total volume of 15 µl (12 µl PCR mix and 3 µl template DNA) was produced. For every sample, the PCR mix included forward primer 1.5 µl, reverse primer 1.5 µl, PCR master mix 7.5 µl and nucleasefree water 1.5 µl. The PCR conditions were as follows, initial denaturation at 95°C for 2 minutes, denaturation at 94°C for 20 seconds, annealing at 54°C for 20 seconds, extension at 72° for 40 seconds, and final extension at 72° for 1 minute. The PCR was run for 30 cycles.

Primer	Sequence $(5'$ -3')	Target	Annealing	Location	Product
			temp $(^{\circ}C)$		size (bp)
PA-GS-F	GACGGGTGAGTAATGCCTA	Pseudomonas species	54	95-113	618
PA-GS-R	CACTGGTGTTCCTTCCTATA			693-712	
PA-SS-F	GGGGGATCTTCGGACCTCA	P. aeruginosa	58	189-206	956
PA-SS-R	TCCTTAGAGTGCCCACCCG			1124-1144	

Table 1: 16s rDNA*-*based primers and their characteristics. Reference: Spilker et al., 2004

2.8 FTIR analysis

Fourier Transform Infrared Spectroscopy (FTIR) was done using a spectrophotometer (IRPrestige-21, SHIMADZU, Japan). This was used to analyze how functional groups were changed upon *Z. atratus* larval consumption (Peng et al., 2020). This analysis was done on fresh LDPE, LLDPE, and EPS (control) as well as after these were consumed as feed, and larvae frass from all the systems.

Firstly the samples were mixed with 30 % hydrogen peroxide (H_2O_2) at 60 °C and treated for 3 days to dissolve contaminants. Then the H_2O_2 was removed and the samples were washed with distilled water. After that, 5.3 mol/l of aqueous sodium iodide (NaI) was added to the samples, incubated for 7 hours, and then gravity separation was conducted. The samples were air-dried, and the FTIR machine was used to analyze them. The resolution of analysis was 4 cm⁻¹, and accumulation was done 45 times at a wavelength range of $4000 - 600$ cm⁻¹.

2.9 SEM Analysis

Scanning Electron Microscopy (SEM) was performed as previously described (Taghavi et al., 2021) using an ultra-high-resolution Schottky Field Emission Scanning Electron Microscope (JSM-7610F, JEOL, Japan). SEM images were taken after LDPE, LLDPE, and EPS samples were incubated in MS broth with isolated bacteria for two months. Untreated LDPE, LLDPE, and EPS samples were used as controls.

After 2 months of incubation with isolated bacteria, plastics were cut by 1 cm x 1 cm. They were placed in a glass vial and air dried at 50°C for 2 days. Then the fully dried samples were mounted on a suitable specimen stub with carbon tape and coated the sample with a thin layer of platinum using a sputter coater. Now the samples were ready to mount in the SEM machine. For LDPE and LLDPE, the SEM image was captured in 1µm length with 3000x magnification. For EPS, it was captured in 10µm length with 1000x magnification. As different plastic type has different structures, different magnifications and lengths have been adjusted accordingly.

2.10 Whole Genome Sequencing, Assembly, and Species Identification

The confirmed samples for *Pseudomonas* were sent to the International Centre for Diarrhoeal Disease Research, Bangladesh (icddr, b) for Whole Genome Sequencing (WGS). The genomic DNA from the *Pseudomonas* samples was extracted using the Wizard® Genomic DNA Purification Kit (Promega, WI, USA) according to the manufacturer's instructions. WGS of the extracted DNA was performed using the NextSeq 550 Sequencer (Illumina, CA, USA). Illumina DNA prep library kit was used to prepare the genomic fragments before carrying out sequencing with a paired-end layout of 150 bp. Generated paired-end reads were trimmed to remove adapters using DRAGEN Bio-IT Platform. The generated trimmed reads were checked for quality using FastQC (LaMar, 2015). After that, SPAdes 3.15.3 under the Galaxy server was used to assemble the reads (Bankevich et al., 2012). Since the sequencing depth was higher than 50X for both samples, --isolate option was enabled to generate both contigs and scaffolds. For the identification of species and strains, the assembled genomes were uploaded to PubMLST "Identify Species" tool and searched (Jolley et al., 2018). Furthermore, the annotated 16S ribosomal RNA (rRNA) sequences from the genomes were compared with known 16S rRNA sequences using NCBI BLASTN (Altschul et al., 1990). Finally, PA-SS primers were used to carry out species-specific PCR for *P. aeruginosa*, and the amplicons were analyzed in agarose gel electrophoresis (Spilker et al., 2004).

In the PCR process, the primer pair used was PA-SS-F and PA-SS-R (Table 1). A total volume of 15 µl (12 µl PCR mix and 3 µl template DNA) was produced. For every sample, the PCR mix included forward primer 1.5 µl, reverse primer 1.5 µl, PCR master mix 7.5 µl and nucleasefree water 1.5 µl. The PCR conditions were as follows, initial denaturation at 95°C for 2 minutes, denaturation at 94°C for 20 seconds, annealing at 58°C for 20 seconds, extension at 72° for 40 seconds, and final extension at 72° for 1 minute. The PCR was run for 30 cycles.

To determine the MLST (Multilocus Sequence Typing) type, the assembled genomes were uploaded to the PubMLST *P. aeruginosa* typing database for analyzing MLST loci. The typing is based on an allelic profile composed of 7 separate loci: *acsA, aroE, quaA, mutL, nuoD, ppsA*, and *trpE*. The assembly quality of the contigs and scaffolds was analyzed using QUAST 5.2.0 (Gurevich et al., 2013). Reference genome FASTA file and GFF (General Feature Format) file for *P. aeruginosa* were provided for improved assessment. Assembled genomes were analyzed with the MOB-Recon 3.0.3 under the Galaxy server to identify possible plasmid sequences (Robertson & Nash, 2018). Matched plasmid sequences were retrieved and compared with the assembled sequences using BLASTN. The assemblies were submitted to NCBI SRA (Sequence Read Archive).

2.11 Genome Annotation and Analysis

For annotation, the assembled genomes were annotated using Prokka 1.14.6 under the Galaxy Europe server, with species specified as *P. aeruginosa* and genus-specific BLAST database enabled (Seemann, 2014). To check for genome completeness, the assembled genomes were compared against a set of 1617 marker genes using the CheckM taxonomy workflow under the Galaxy Europe server (Parks et al., 2015). The assembled genome and the associated GFF3 files were provided to Operon-mapper to predict potential operons throughout the genomes (Taboada et al., 2018). Proteins were also uploaded to the BLASTKoala server to assign K numbers to the proteins (Kanehisa et al., 2016).

Moreover, to identify potential plastic-degrading enzymes among the annotated proteins, the protein sequences predicted from each genome were uploaded to the PlasticDB "Annotate genome" tool for BLASTP analysis against sequences of known plastic-degrading enzymes (Gambarini et al., 2022). Additionally, to find possible antibiotic resistance genes in the genomes, the assembled genomes were uploaded to CARD-RGI (Comprehensive Antibiotic Resistance Database - Resistance Gene Identifier) to find matches with known antibiotic resistance genes (Alcock et al., 2023). The comparison looked for perfect matches as well as strict matches with over 95% identity. Lastly, to identify possible virulence factors, the protein sequences predicted from each genome were uploaded to VFanalyzer; the genus was specified as *Pseudomonas* (Chen et al., 2005).

For pan-genome analysis, the RefSeq genome assemblies of the following *P. aeruginosa* strains were retrieved from NCBI: biosurfactant producing strains UCBPP-PA14 and PA7; PE degrading strain E7; LDPE degrading strain PA01; highly virulent strains PA14_ASM2549037, PA14_ASM2549039, PA14_ASM2549041, and PA14_ASM2549047; and BWHPSA013, which has the same MLST type as the assembled genomes (Gutiérrez-Gómez et al., 2019; Jeon & Kim, 2015; Kyaw et al., 2012; Lee et al., 2006; Toribio et al., 2010). The genomes were annotated with Prokka as before and then compared with the annotated genomes using Roary 3.13.0 under Galaxy Europe server, with BLASTP cutoff set at 99% (Page et al., 2015). For further confirmation, proteins of interest were further analyzed at InterProScan for identification of motifs and domains (Paysan-Lafosse et al., 2023).

16S rRNA sequences that showed more than 97% sequence identity with sequences from both isolates were retrieved from NCBI. The 16S rRNA sequences from the Prokka-annotated *P. aeruginosa* strains were also retrieved. All 16S rRNA sequences were submitted to the NGPhylogeny web service for phylogenetic tree construction (Lemoine et al., 2019). The sequences were aligned using MAFFT, followed by alignment curation using BMGE, tree inference using PhyML+SMS, and finally, tree rendering using Newick display (Criscuolo & Gribaldo, 2010; Guindon et al., 2010; Junier & Zdobnov, 2010; Katoh & Standley, 2013; Lefort et al., 2017; Lemoine et al., 2018). The rendered tree was visualized and modified in Interactive Tree Of Life (iTOL) (Letunic & Bork, 2021). Additionally, a core gene alignment was generated by Roary, which was analyzed through a similar pipeline to render and visualize a phylogenetic tree of the strains.

2.12 Bio-surfactant Assay

Biosurfactant assay were done to see whether the isolated bacteria could produce any biosurfactant that can be used during oil spill bioremediation. In this assay, paraffin oil was used. This assay had two tests: the oil spread test and the drop collapse test. For these tests, 48 H cultures were produced. In addition to these cultures, a 4-month-old bacteria culture (using nutrient broth with LDPE, LLDPE, and EPS) and 1 Year 1-year-old bacteria culture (using MSB media with plastics as sole carbon source) were tested.

Now 9 systems were designed,

i. Nutrient broth inoculated with bacteria

- ii. Nutrient broth inoculated with bacteria and added soybean oil
- iii. Nutrient broth inoculated with bacteria added paraffin oil
- iv. Nutrient broth (Negative control)
- v. Nutrient broth (Negative control)
- vi. 4 month old bacteria culture+ Nutrient broth+ soybean oil
- vii. 4 month old bacteria culture+ Nutrient broth+ paraffin oil

viii. 4-month-old bacteria culture

ix. 1-Year-old bacteria culture

Here, soybean oil and paraffin oil in the culture tubes increase stress on the inoculated bacteria. It is hoped that they will produce more biosurfactants due to the environmental stress.

2.12.1 Oil Spread Test

An oil spread test was done with all the systems. First, 1 ml of cultures from all the systems were collected and centrifuged at 5000 rpm for 10 minutes. Culture supernatant was collected. Next, in a petri dish, 20ml water was poured. In the water, 2 ml of paraffin oil was added. Now, 1ml culture supernatant was added on top of the oil at the center of the petri dish. If the oil produced zone, it meant the culture supernatant contained biosurfactant whereas no zone means no biosurfactant. As a positive control, 10% Triton X-100 was used.

2.12.2 Drop Collapse Test

A drop collapse test was done with all the systems. In a vial, 1 ml of paraffin oil was taken. Then, 500µl culture was added. The result was observed after 1 minute. If the culture drop was flat or seemed like it was mixing with oil, then it was a positive result. If the culture drop was round with a clear shape, then it was a negative result. Here, as a positive control, 10% Triton X-100 was used.

2.13 Antimicrobial Susceptibility Test

Antimicrobial susceptibility tests were performed from the WGS data. According to the WGS data for AMR genes, antibiotics from different classes were chosen. Firstly, isolated bacteria were grown on nutrient agar media. Mueller–Hinton agar was used for AST. With 24H culture, 0.5 McFarland standard solutions were prepared. Using a sterile cotton swab, bacteria were inoculated with the lawn method. AST was done using the disk diffusion procedure. So, the disks of the selected antibiotics were put on the surface of the media. The culture incubation period was 24 hours at 37° C. After incubation, the zone was measured (diameter in mm) according to CLSI guidelines.

2.14 Statistical Analysis

The analysis of plastic consumption and survivability were assessed by one-way ANOVA coupled with Tukey's honest significant difference (HSD) post-hoc test. The analysis was performed via GraphPad Prism version 8 (GraphPad Software, MA, USA).

Chapter 3

Results

3.1 *Zophobas atratus* **Plastic Consumption Rate and Survival Rate**

The larvae excreted brown particle-like frass. The frass contained a trace amount of plastics, however as the larvae can recycle the frass, as time passed, the amount of plastic particles also decreased. On another note, larvae showed cannibalism. Though in most of the system, dead larvae were removed as soon as possible, some started cannibalization. On the other hand, molting was seen in all systems (Figure 1). Among all the systems, only two larvae became pupae, and they successfully became adult insects after isolating them from the rest of the group (Figure 1). Of the two adults, one was male and one was female (Figure 1).

The larvae consumed all three types of plastics (Figure 4). However, the degree of consumption is different. PC for LLDPE was on average 15.12% in 36 days, with 0.41 mg 20 larvae⁻¹ day⁻¹ (Figure 5). The SR for this group was found to be $87\% \pm 10.4\%$ (Figure 6). Similarly, the PC for the larvae feeding on LDPE was on average 24.04% in 36 days with 7.37 mg 20 larvae-1 day⁻¹ with 85% \pm 10% SR (Figure 5, 6). For EPS, the PC was on average 20.01% in 36 days with 12.39 mg 20 larvae⁻¹ day⁻¹ (Figure 5). The SR for this group was found to be $90\% \pm 8.66\%$ (Figure 6). In PCN, the SR was 100%, whereas NCN had $68\% \pm 7.63\%$ (Figure 6). For all the groups except the PCN, the LW decreased over time (Figure 3). When compared with PCN LW, significant change can be observed from 36 days onwards for LDPE and EPS and from 24 days onwards for LLDPE (Table 2). However, there is no significant difference in LW decrease among the three plastic types.

Figure 1: The larvae lifecycle observed during experimentation. (1) Larvae. (2) Larvae after molting. (3) Prepupae stage. (4) Pupae stage. (5) Adult beetle. (6) Male and female beetle

Figure 2: Larvae frass. The larvae excreted brown particle like frass

Days	PCN LW(g)	NCN LW(g)	LLDPE LW (g)	LDPE	EPS LW (g)
				LW(g)	
$\mathbf{0}$	0.392 ± 0.017	0.387 ± 0.023	0.367 ± 0.013	0.392 ± 0.017	0.413 ± 0.0066
21	0.434 ± 0.054	$0.35 + 0.028$	0.345 ± 0.039	$0.362 + 0.040$	0.355 ± 0.0385
24	0.443 ± 0.057	0.339 ± 0.025	$0.344 \pm 0.049*$	0.354 ± 0.036	0.357 ± 0.036
36	0.466 ± 0.061	0.318 ± 0.016	$0.335 \pm 0.029*$	$0.329 \pm 0.048*$	$0.339 \pm 0.0370*$
41	0.480 ± 0.055	0.304 ± 0.017	$0.332 \pm 0.057*$	$0.317 + 0.054*$	$0.326 \pm 0.045*$
51	0.516 ± 0.043	$0.273 + 0.016$	$0.331 \pm 0.058*$	$0.303 + 0.069*$	$0.319 + 0.043*$

Table 2: Larvae Weight measurement for 51 days (average of triplicated results).

 $*$ = significant value (p<0.05) when compared with PCN LW. No significant LW change when compared with NCN LW. SR of NCN was the lowest among the tests, which implies plastic consumption gave enough nutrients to the larvae to survive more. PCN=positive control, NCN=negative control, LW= larvae average weight, LLDPE= linear low density polyethylene, LDPE=low density polyethylene, EPS=expanded polystyrene

Figure 3: Time analysis of LW. When compared with PCN LW, significant (p < 0.05) change can be observed from 36 days onwards for LDPE and EPS and from 24 days onwards for LLDPE. No significant difference in LW decrease among LDPE, EPS, and LLDPE

Figure 4: Larvae consumed all 3 types of plastic. (A) LDPE consumption by the larvae in 36 days, (i) control LDPE, (ii) after 15 days, (iii) after 36 days. (B) LLDPE consumption by the larvae in 36 days, (i) control LLDPE, (ii) after 36 days, (iii) close-up shot in for a better view. (C) EPS consumption by the larvae in 51 days, (i) control EPS, (ii) after 36 days, (iii) after 51 days. LLDPE=linear low-density polyethylene, LDPE=low density polyethylene, EPS=expanded polystyrene

Figure 5: The larvae consumed the plastics at various rates. (A) PC with different plastic types (triplicated result) where LDPE has the highest rate and LLDPE has the lowest. (B) PC with a time analysis of 36 days. The points represent the mean \pm SEM (n = 3). Data were analyzed using one-way ANOVA coupled with Tukey's honest significant difference (HSD) posthoc test

Figure 6: Larvae survivability showed EPS as the best plastic type. (A) SR%. PCN had 100 % SR and NCN had 68 %. Among the plastic types, EPS had the highest SR whereas LDPE had the lowest. (B) Time analysis of SR% in 36 days. In all the statistical analyses, the points represent the mean \pm SEM (n = 3). Data were analyzed using one-way ANOVA coupled with Tukey's honest significant difference (HSD) post-hoc test. SR = survival rate, PCN = positive control, NCN = negative control, LW = larvae average weight, LLDPE = linear lowdensity polyethylene, LDPE = low density polyethylene, EPS = expanded polystyrene

3.2 FTIR Analysis

FTIR analysis showed proof of oxidation in the gut. As stated earlier, fresh plastic (as a control), larvae consumed plastic, and larvae frass were analyzed for each system. New functional groups at 1075-1150 cm⁻¹ (-C-O stretch), 1700 cm⁻¹ (-C=O stretch), and 3440 cm⁻¹ (Re-OH stretch) wavenumbers were identified (Figure 7). For all three plastic types, these three functional groups were only found in the frass whereas control and consumed plastics had almost similar results. In the control and consumed plastics, no oxygen bonded with another element was identified. Additionally for LDPE, 1700 cm^{-1} (-C=O stretch) peak was also found in the consumed part which was absent in the control.

Figure 7: FTIR analysis where three new 1075-1150 cm⁻¹ (-C-O stretch), 1700 cm⁻¹ (-C=O stretch), and 3440 cm-1 (Re-OH stretch) functional groups were found which proves oxidation. (A) Analysis of LDPE. Consumed LDPE had $-C=O$ stretch and frass had all three but control had none of the functional groups of those three. (B) Analysis of LLDPE. Control and consumed had similar results whereas frass had three new functional groups. (C) Analysis of EPS, control, and consumed had none of those three functional groups but frass had.

LLDPE=linear low-density polyethylene, LDPE=low density polyethylene, EPS=expanded polystyrene

3.3 Bacteria Isolation and Identification

Growth was observed after inoculation in MSB from every system every 7 days. After 7 days, no growth was observed for NCB. For PCB, no growth was observed after 40 days. After 60 days of incubation, spread plating was done on selective media, where NA, MacConkey agar, and Cetrimide agar showed bacterial growth. As the target bacteria was *Pseudomonas*, growth on cetrimide agar plate was chosen. Isolated bacteria from LDPE/LLDPE gave a green pigmentation without UV whereas bacteria from EPS gave no color (Figure 8). However, both isolates exhibited florescence under UV. Due to similar types of plastic and morphology, isolates grown on LDPE/LLDPE were designated as "PDB-1" and the isolate grown on EPS was designated as "PDB-2". These two isolates were confirmed *Pseudomonas* with PCR which gave its characteristic band in gel electrophoresis.

Figure 8: Isolated bacteria characteristics. Bacteria from LDPE and LLDPE media displayed green pigmentation without UV, while those from EPS media lacked this pigmentation. All exhibited fluorescence under UV, a positive *Pseudomonas* characteristic. Isolates PDB-1 (from LDPE and LLDPE) and PDB-2 (from EPS) demonstrated growth in MS medium with specific plastics as the sole carbon source, indicating plasticdegrading biochemical pathways. In the figure (A) green pigmentation for PDB-1, (B) no green pigmentation for PDB-2.

3.4 SEM Analysis

Both PDB-1 and PDB-2 had shown surface degradation as seen using SEM. In this analysis, LDPE and LLDPE were incubated with PDB-1 and EPS with PDB-2 in MSB broth where the plastics were the sole carbon source. After incubation, all incubated plastics showed signs of surface degradation when compared with controls. The control samples exhibited smooth surfaces, while the incubated samples displayed rough, fragmented surfaces characterized by the presence of bacteria and biofilms. In the case of LDPE, control samples showed minor irregularities with isolated microplastic particles, whereas the incubated LDPE exhibited surface erosion, featuring significant pits, bumps, microplastic fragments, and biofilm formation. Though images were taken in different lengths with different magnifications, for consistency and to understand better, the image that was captured in 1µm length with 3,000x magnification has been added here (Figure 9). Similarly, LLDPE control samples demonstrated smooth surfaces, contrasting with the rough, bumpy surfaces of the incubated samples, accompanied by bacterial colonization and biofilm formation. Though images were taken in different lengths with different magnifications, for consistency and to understand better, the image that was captured in 1µm length with 10,000x magnification has been added here (Figure 9). Likewise, EPS control samples displayed smooth surfaces, whereas the incubated EPS exhibited tears, holes, and biofilm formation at various locations. Though images were taken in different lengths with different magnifications, for consistency and to understand better, the image that was captured in 10µm length with 1,000x magnification has been added here (Figure 9).

Figure 9: SEM analysis of LDPE, LLDPE and EPS proved degradation by the two isolated bacteria. (A) Control and degraded LDPE by PDB-1 isolate. Control had a comparatively smooth surface with little microplastics whereas degraded LDPE had a rough surface with many microplastics. The image was captured in 1µm length with 3000x magnification. (B) Control and degraded LLDPE by PDB-1 isolate. Control had a smooth surface whereas degraded LLDPE had a rough surface with bacteria and biofilms. The image was captured in 1µm length with 10000x magnification. (C) Control and degraded EPS by PDB-2 isolate. Control had a smooth surface whereas degraded EPS had a rough surface with tears, holes, and biofilms. The image was captured in 10µm length with 1000x magnification. LLDPE= linear low-density polyethylene, LDPE=low density polyethylene, EPS=expanded polystyrene

3.5 Genome Assembly and Identification

According to the sequencing data, PDB-1 and PDB-2 were sequenced with coverages of 298.7 and 142.1, respectively. Upon assembly, PDB-1 yielded 2254 contigs and 2164 scaffold assemblies, with 152 contigs and 62 scaffolds exceeding 500 bp in length. Conversely, the PDB-2 assembly generated 1218 contigs and 1133 scaffold assemblies, with 140 contigs and 55 scaffolds over 500 bp in length. The NG50 value for PDB-1 contigs was 81867 bps, increasing to 246604 bps for the scaffolds. Similarly, PDB-2 contigs had an NG50 value of 87657 bps, rising to 258807 bps for the scaffolds. Scaffolds exhibited slightly higher genome coverage than contigs in both cases, thus chosen as the genomic sequence. The assemblies' quality was indicated by the low contig count (<100) and high NG50 value (>50000bp). The assemblies are available in NCBI under **BioProject ID PRJNA1005894**.

PubMLST identified the genome as *P. aeruginosa* with 100% confidence, showing 55 exact matches with known sequences. NCBI BLASTN revealed nearly identical 16S rRNA sequences to known *P. aeruginosa* sequences, and as both isolates have identical 16s rRNA sequences, it suggests a close phylogenetic relationship. Additionally, PCR with PA-SS primer pairs confirmed their identity as *P. aeruginosa* (Figure 10). Furthermore, MLST typing classified the genomes as ST170 (Loci numbers - acsA:36, aroE:5, quaA:29, mutL:7, nuoD:4, ppsA:10, and trpE:7).

Lastly, MOB-Recon analysis detected probable plasmid sequences in both genomes, matching with known *P. aeruginosa* plasmid sequences. Plasmid pHOU1-1 from *P. aeruginosa* strain HOU1 matched scaffold sequences from both genomes. Comparison with the plasmid sequence (GenBank Accession: CP042268) identified 2 scaffold sequences from PDB-1 and 3 scaffold sequences from PDB-2 as probable plasmid sequences.

Figure 10: Isolated bacteria characteristics. Gel electrophoresis using PA-SS primer set with characteristic DNA band at 956 bp. (+)-ve control had already identified *P. aeruginosa* whereas (-)-ve control had no sample DNA. All the bands for PDB-1 and PDB-2 aligned at 956 bp along with the (+)-ve control which is the characteristic of *P. aeruginosa* . No band at (-)-ve control means there was no contamination during the experimentation. The DNA ladder used during gel electrophoresis was AMPIGENE® DNA Ladder 50-1,500 bp.

PDB-1 and PDB-2 16s rDNA sequences are given below:

>PDB1_16S ribosomal RNA

TGAAGAGTTTGATCATGGCTCAGATTGAACGCTGGCGGCAGGCCTAACACATGCAAGTCG AGCGGATGAAGGGAGCTTGCTCCTGGATTCAGCGGCGGACGGGTGAGTAATGCCTAGGAA TCTGCCTGGTAGTGGGGGATAACGTCCGGAAACGGGCGCTAATACCGCATACGTCCTGAG GGAGAAAGTGGGGGATCTTCGGACCTCACGCTATCAGATGAGCCTAGGTCGGATTAGCTA GTTGGTGGGGTAAAGGCCTACCAAGGCGACGATCCGTAACTGGTCTGAGAGGATGATCAG TCACACTGGAACTGAGACACGGTCCAGACTCCTACGGGAGGCAGCAGTGGGGAATATTGG ACAATGGGCGAAAGCCTGATCCAGCCATGCCGCGTGTGTGAAGAAGGTCTTCGGATTGTA AAGCACTTTAAGTTGGGAGGAAGGGCAGTAAGTTAATACCTTGCTGTTTTGACGTTACCA ACAGAATAAGCACCGGCTAACTTCGTGCCAGCAGCCGCGGTAATACGAAGGGTGCAAGCG TTAATCGGAATTACTGGGCGTAAAGCGCGCGTAGGTGGTTCAGCAAGTTGGATGTGAAAT CCCCGGGCTCAACCTGGGAACTGCATCCAAAACTACTGAGCTAGAGTACGGTAGAGGGTG GTGGAATTTCCTGTGTAGCGGTGAAATGCGTAGATATAGGAAGGAACACCAGTGGCGAAG GCGACCACCTGGACTGATACTGACACTGAGGTGCGAAAGCGTGGGGAGCAAACAGGATTA GATACCCTGGTAGTCCACGCCGTAAACGATGTCGACTAGCCGTTGGGATCCTTGAGATCT TAGTGGCGCAGCTAACGCGATAAGTCGACCGCCTGGGGAGTACGGCCGCAAGGTTAAAAC TCAAATGAATTGACGGGGGCCCGCACAAGCGGTGGAGCATGTGGTTTAATTCGAAGCAAC GCGAAGAACCTTACCTGGCCTTGACATGCTGAGAACTTTCCAGAGATGGATTGGTGCCTT CGGGAACTCAGACACAGGTGCTGCATGGCTGTCGTCAGCTCGTGTCGTGAGATGTTGGGT TAAGTCCCGTAACGAGCGCAACCCTTGTCCTTAGTTACCAGCACCTCGGGTGGGCACTCT AAGGAGACTGCCGGTGACAAACCGGAGGAAGGTGGGGATGACGTCAAGTCATCATGGCCC TTACGGCCAGGGCTACACACGTGCTACAATGGTCGGTACAAAGGGTTGCCAAGCCGCGAG GTGGAGCTAATCCCATAAAACCGATCGTAGTCCGGATCGCAGTCTGCAACTCGACTGCGT GAAGTCGGAATCGCTAGTAATCGTGAATCAGAATGTCACGGTGAATACGTTCCCGGGCCT TGTACACACCGCCCGTCACACCATGGGAGTGGGTTGCTCCAGAAGTAGCTAGTCTAACCG CAAGGGGGACGGTTACCACGGAGTGATTCATGACTGGGGTGAAGTCGTAACAAGGTAGCC GTAGGGGAACCTGCGGCTGGATCACCTCCTT

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>PDB2_16S ribosomal RNA

TGAAGAGTTTGATCATGGCTCAGATTGAACGCTGGCGGCAGGCCTAACACATGCAAGTCG AGCGGATGAAGGGAGCTTGCTCCTGGATTCAGCGGCGGACGGGTGAGTAATGCCTAGGAA TCTGCCTGGTAGTGGGGGATAACGTCCGGAAACGGGCGCTAATACCGCATACGTCCTGAG GGAGAAAGTGGGGGATCTTCGGACCTCACGCTATCAGATGAGCCTAGGTCGGATTAGCTA GTTGGTGGGGTAAAGGCCTACCAAGGCGACGATCCGTAACTGGTCTGAGAGGATGATCAG TCACACTGGAACTGAGACACGGTCCAGACTCCTACGGGAGGCAGCAGTGGGGAATATTGG ACAATGGGCGAAAGCCTGATCCAGCCATGCCGCGTGTGTGAAGAAGGTCTTCGGATTGTA AAGCACTTTAAGTTGGGAGGAAGGGCAGTAAGTTAATACCTTGCTGTTTTGACGTTACCA ACAGAATAAGCACCGGCTAACTTCGTGCCAGCAGCCGCGGTAATACGAAGGGTGCAAGCG TTAATCGGAATTACTGGGCGTAAAGCGCGCGTAGGTGGTTCAGCAAGTTGGATGTGAAAT CCCCGGGCTCAACCTGGGAACTGCATCCAAAACTACTGAGCTAGAGTACGGTAGAGGGTG GTGGAATTTCCTGTGTAGCGGTGAAATGCGTAGATATAGGAAGGAACACCAGTGGCGAAG GCGACCACCTGGACTGATACTGACACTGAGGTGCGAAAGCGTGGGGAGCAAACAGGATTA GATACCCTGGTAGTCCACGCCGTAAACGATGTCGACTAGCCGTTGGGATCCTTGAGATCT TAGTGGCGCAGCTAACGCGATAAGTCGACCGCCTGGGGAGTACGGCCGCAAGGTTAAAAC TCAAATGAATTGACGGGGGCCCGCACAAGCGGTGGAGCATGTGGTTTAATTCGAAGCAAC GCGAAGAACCTTACCTGGCCTTGACATGCTGAGAACTTTCCAGAGATGGATTGGTGCCTT CGGGAACTCAGACACAGGTGCTGCATGGCTGTCGTCAGCTCGTGTCGTGAGATGTTGGGT TAAGTCCCGTAACGAGCGCAACCCTTGTCCTTAGTTACCAGCACCTCGGGTGGGCACTCT AAGGAGACTGCCGGTGACAAACCGGAGGAAGGTGGGGATGACGTCAAGTCATCATGGCCC TTACGGCCAGGGCTACACACGTGCTACAATGGTCGGTACAAAGGGTTGCCAAGCCGCGAG GTGGAGCTAATCCCATAAAACCGATCGTAGTCCGGATCGCAGTCTGCAACTCGACTGCGT GAAGTCGGAATCGCTAGTAATCGTGAATCAGAATGTCACGGTGAATACGTTCCCGGGCCT TGTACACACCGCCCGTCACACCATGGGAGTGGGTTGCTCCAGAAGTAGCTAGTCTAACCG CAAGGGGGACGGTTACCACGGAGTGATTCATGACTGGGGTGAAGTCGTAACAAGGTAGCC GTAGGGGAACCTGCGGCTGGATCACCTCCTT

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3.6 Annotation of Significant Genes

During the annotation process, PDB-1 and PDB-2 revealed 6011 and 5949 putative genes, respectively, with 5859 and 5804 identified as protein-coding sequences. Moreover, out of the 1617 marker genes in CheckM, 1610 and 1611 were detected in PDB-1 and PDB-2, respectively, indicating completeness levels of 99.64% and 99.66%, respectively. Annotations were provided alongside the assembly at NCBI. BLASTKoala analysis assigned K numbers to 3365 and 3359 proteins from PDB-1 and PDB-2, respectively. When predicted proteins were compared with known plastic-degrading proteins in PlasticDB, four predicted proteins from PDB-1 matched PE-degrading enzymes. Specifically, PDB1_00173 and PDB1_01428 matched with *Psychrobacter* sp. laccase (PlasticDB Protein ID: 00180), while PDB1_00445 and PDB1_02915 matched with *Pseudomonas* sp. alkane hydroxylase (PlasticDB Protein ID: 00061), with the latter two also matching LDPE-degrading enzyme alkane monooxygenase from *Paenibacillus* sp. (PlasticDB Protein ID: 00104). No proteins from PDB-2 matched with known PS-degrading enzymes. In a comparison of genes across strains, 1822 core genes were identified as conserved in all strains, alongside 3232 soft core genes, 1758 shell genes, and 5563 cloud genes, which show varying degrees of conservation across different strains.

Phylogenetic analysis based on 16S rRNA sequences indicated that both PDB-1 and PDB-2 isolates formed close clusters with other strains of *P. aeruginosa*. Additionally, strains of various other *Pseudomonas* species and some species from the recently suggested *Stutzerimonas* genus exhibited high sequence similarity. A more robust phylogenetic tree was constructed using core gene alignment from pan-genome analysis, revealing that PDB-1 and PDB-2 shared the highest similarity with the BWHPSA013 strain, which belongs to the same MLST type (Figure 11).

Among the genes in PDB-1, PDB1 05091 was annotated as the Cytochrome P450 107B1 gene, with P450-driven monooxygenase activity essential for the degradation pathway. Additionally,

genes PDB1_01395, PDB1_01396, and PDB1_01397, annotated as short-chain dehydrogenase, esterase, and Baeyer-Villiger monooxygenase, respectively, were predicted to belong to the same operon. Another operon included PDB1 00458 and PDB 00459, annotated as short-chain alcohol dehydrogenase and Baeyer-Villiger monooxygenase, respectively. Furthermore, PDB1_01294, PDB1_01369, and PDB1_03409 were annotated as alcohol dehydrogenases.

PDB1_00173 was annotated as a laccase, supported by PlasticDB. Genes PDB1_00445 and PDB1 02915 were predicted to encode this enzyme, corroborated by PlasticDB. Comparison with genes from the PE-degrading strain E7, LDPE-degrading strain PA01, and PDB-2 revealed orthologs in most cases with over 99% sequence similarity. However, PDB1_01395 and PDB1_01294 did not have orthologs in E7. PDB1_01395 matched only with orthologs from PDB-2, PA01, and BWHPSA013, while PDB1_01294 matched only with orthologs from PDB-2 and PA01. The annotation of the PA7 strain identified a dye-decolorizing peroxidase (yfeX), corresponding to the protein PSPA7_2468 (GenBank Accession: ABR83505), previously annotated as a hypothetical protein.

The gene PDB2 04415 in PDB-2 is annotated as phenylacetaldehyde dehydrogenase (styD), with highly similar orthologs in all other strains except PA7. Conversely, PDB2 05770, also annotated as phenylacetaldehyde dehydrogenase, lacks such orthologs, but the encoded protein is too short to be independently functional. Both PDB2_04415 and PDB2_05770 proteins match with Streptomyces sp. PEG aldehyde dehydrogenase from PlasticDB (PlasticDB Protein ID: 00035). Among PDB-2 genes, PDB2_03471 is annotated as the Cytochrome P450 107B1 gene, orthologous to PDB1_05091. Additionally, both PDB2_03017 and PDB2_03887 genes are annotated as alkane hydroxylases, with their proteins matching with *Pseudomonas* sp. alkane hydroxylase and *Paenibacillus* sp. alkane monooxygenase in PlasticDB.

In PDB-1, RGI identified 25 perfect matches with antibiotic resistance genes, while in PDB-2, 24 perfect matches were found. Most resistance was against fluoroquinolone-, tetracycline-, and phenicol-class antibiotics, with antibiotic efflux being the most common resistance mechanism (Table 3).

Virulence factors present in the isolates were analyzed and compared with other strains. Different genes for adherence, flagella, antiphagocytic activity, etc have been found in both isolates and an important pathway involved in the production of rhamnolipid, a biosurfactant, was identified with three genes, rhlA, rhlB, and rhlC (Table 4).

SI	Isolate	Protein ID	Best Hit AR	Drug Class	Resistance	AMR Gene Family
			$\mathbf 0$		Mechanism	
$\mathbf{1}$	$PDB-1$	PDB1_0008	MexL	macrolide	antibiotic	resistance-
		1		antibiotic;	efflux	nodulation-cell
		hypothetical		tetracycline		division (RND)
		protein		antibiotic;		antibiotic efflux
				disinfecting agents		pump
				and antiseptics		
$\overline{2}$	$PDB-1$	PDB1_0022	YajC	fluoroquinolone	antibiotic	resistance-
		9 Sec		antibiotic;	efflux	nodulation-cell
		translocon		cephalosporin;		division (RND)
		accessory		glycylcycline;		efflux antibiotic
		complex		penam;		pump
		subunit YajC		tetracycline		
				antibiotic;		
				oxazolidinone		
				antibiotic;		
				glycopeptide		
				antibiotic;		
				rifamycin		
				antibiotic;		
				phenicol		
				antibiotic;		

Table 3: Genes in isolates PDB-1 and PDB-2 showing perfect matches with known Antimicrobial Resistance (AMR) genes, as identified by CARD RGI

Table 4: Virulence Factors identified in isolates PDB-1 and PDB-2, as determined by Vfanalyzer

Figure 11: Phylogenetic analysis of the isolated strains. (A) Tree built from 16S rRNA sequences of the isolates PDB1 and PDB2, other *P. aeruginosa* strains, and closely aligned 16S rRNA sequences. (B) Tree built from the core gene alignment among PDB1, PDB2, and other *P. aeruginosa* strains used for the pan-genome analysis

3.7 Biosurfactant Assay

Both the oil spread test and drop collapse test showed positive biosurfactant activity. For the oil spread test, all the systems except negative control showed dispersion upon interaction. When mixed, the systems with biosurfactant produced oil-water emulsion comparable to positive control. (Figure 12)

On the other hand, for the drop collapse test, negative control and system 9 which was the 1 year-old bacteria in MSB had a round-shaped drop. However, system 9 gave mixed results as one part was flat and one part was round shape (Figure 13)

Figure 12: Oil spread test. As a continuous steps, it was shown when cell free supernatant interacts with oil, it disperses the oil and make a zone proving biosurfactant activity

Figure 13: Drop collapse test. 1. Nutrient broth inoculated with bacteria. 2. Nutrient broth inoculated with bacteria and added soybean oil. 3. Nutrient broth inoculated with bacteria added paraffin oil. 4. Nutrient broth (Negative control). 6. 4-month-old bacteria culture+ Nutrient broth+ soybean oil. 7. 4-month-old bacteria culture+ Nutrient broth+ paraffin oil. 8. 4-month-old bacteria culture. 9. 1-Year-old bacteria culture with mixed result

3.8 AST Analysis

According to the perfect matches with antibiotic-resistant genes, 7 different classes of antibiotics were chosen. Tetracycline 30 mg (tetracycline class), ampicillin 10mg (penicillin class), erythromycin 15 mg (macrolide class), ceftazidime 30 mcg (third-generation cephalosporin), tigercycline 15 mcg (glycylcycline class), kanamycin 5mcg (aminoglycoside class) and Meropenem+ EDTA (carbapenem class) were chosen as representatives.

Both PDB-1 and PDB-2 were found to be resistant to all the listed antibiotics. For PDB-1, a small zone can be observed for tetracycline, kanamycin, and tigercycline though they were all less than the required zone diameter for them to be susceptible. The same result was observed for PDB-2. (Figure)

Figure 14: AST analysis of PDB-1 and PDB-2. Both bacteria were resistant to seven class of antibiotics as seen from WGS analysis

Chapter 4

Discussion

This study has proved the biodegradation capability of *Zophobas atratus* larvae and its gut bacteria. In this study, larvae ate 15.12% LLDPE, 24.04% LDPE, and 20.01% EPS in 36 days with respective survival rates of $87\% \pm 10.4\%$, $85\% \pm 10\%$, $90\% \pm 8.66\%$ with brown particlelike frass excreting in every system. This result is comparable to the other studies done on the same larvae, for example (Peng et al., 2020) had almost the same PC and SR for LDPE. the small difference between SR and PC may lie in the absence of any controlled temperature and humidity in this study. For EPS, the outcome is comparable to the study by (Yang et al., 2020). The SR values show that larvae can survive with EPS better than the other two types, but they can consume LDPE at a much higher rate than the others. For the frass and cannibalism, other studies also reported the same results and the frass structure was distinct.

For all the groups except the PCN, the LW decreased over time, similar to the observations of (Luo et al., 2021). When compared with PCN LW, significant change can be observed from 36 days onwards for LDPE and EPS and from 24 days onwards for LLDPE. Though there is no significant difference in LW decrease among the three plastic types, a decrease in weight when compared with SR suggests that the larvae can only survive by eating only plastic, not flourish. This result also confirms why even after molting, LW decreased. The LW values are more dissimilar when compared to other studies. Though cannibalism was observed in the system, the SR for NCN was much lower than the others. So, even if cannibalism occurred, it provided inefficient nutrition which contributed to higher larvae mortality. On the other hand, those larvae who consumed plastic had better nutrition and survived at a higher rate. In the LW graph, NCN consistently showed the lowest value. This signifies the lower growth rate of the larvae when plastic feed was absent. In combination, the lower SR and LW values when plastic feed was absent prove that the polymers provided nutrition alongside cannibalism to the larvae. One of the aims of the experiments was to see how the larvae can degrade plastics when exposed to high temperatures and humidity. This is the expected environmental condition when performing in situ bioremediation in Bangladesh and similar tropical locations. Hence they were not placed in an incubator and had to cope with high $(41^{\circ}C)$ to moderate temperatures with varying humidity. This unique condition may explain the differences in observations with other studies that used insect incubators.

Furthermore, FTIR analysis proved the formation of new functional groups. As stated earlier, for all the plastic types, control plastic, consumed plastic, and frass were analyzed. The data was compared among these three types and the main new functional groups that were identified were at 1075-1150 cm⁻¹ (-C-O stretch), 1700 cm⁻¹ (-C=O stretch), and 3440 cm⁻¹ (Re-OH stretch) wavenumbers, a similar observation to Peng *et al.*, (2020)**.** These functional groups indicate oxidation of the ingested LDPE, LLDPE, and EPS in the larval gut. For all three plastic types, these three functional groups were found in the frass whereas control and consumed plastics had almost similar results. According to previous studies, the incorporation of oxygen functional groups is considered the preliminary and important step towards plastic degradation (Gautam et al., 2007; Shah et al., 2008). Findings of these groups in the frass indicate plastic metabolism/degradation.

As the frass collected was produced by the larvae by eating only plastics, it is evident that some change in the plastic structure was made during consumption. If these larvae could secrete plastic-degrading enzymes like *Galleria mellonella,* FTIR analysis would have found a difference between the control plastic and consumed plastic (Sanluis-Verdes et al., 2022). As for all the plastic types, control and consumed plastics showed nearly the same data, it can be said that the oxidation must be done during digestion. The 1700 cm^{-1} peak in consumed LDPE can be a false positive data. Previous studies confirmed that when the gut microbiome of the larvae was inhibited by antibiotics, the plastic degradation capability was also inhibited (Peng et al., 2020; Yang et al., 2020). As a result, this oxidation is more likely caused by the gut microbiome. The frass had residual plastics, which also means they could not fully degrade the plastic that they consume.

For bacterial growth, as *P. aeruginosa* was the target bacteria, spread plating was carried out after incubation was complete with the mediums, NA, MacConkey agar, and Cetrimide agar. Colony growth on Cetrimide agar was chosen for further analysis, as it is used for the selective isolation of *Pseudomonas,* the target bacteria (Brown & Lowbury, 1965)*.* Using this media also enhances the production of *Pseudomonas* pigments pyocyanin and pyoverdine, which show a characteristic blue-green and yellow-green color respectively. This aided the pigmentation analysis required for differentiation using colony morphology. As bacteria from all three systems (LDPE-, LLDPE-, and EPS-supplemented media) could grow on Cetrimide agar, whereas no bacteria were visible from PCB and NCB, this proves that without a carbon source, those frass bacteria cannot survive for more than 7 days. Even if there is an additional carbon source such as glucose, it is limiting; after glucose gets depleted, bacteria die. On the other hand, as long as plastics are available as carbon sources, the bacteria survive.

From the SEM analysis, it was proven that both PDB-1 and PDB-2 can degrade plastics as they showed surface degradation under SEM when compared with controls. After two months of incubation, all plastics showed surface degradation; control samples had smooth surfaces, whereas rough, broken surfaces with bacteria and biofilms could be seen in the incubated ones. The LDPE results are comparable to a previous study by (Khandare et al., 2021). The degradation pattern of LLDPE is similar to the study by (Shabbir et al., 2020).

The WGS results were analyzed with different *P. aeruginosa* strains along with analyzing differences between PDB-1 and PDb-2. Differentiation between PDB-1 and PDB-2 and comparison with other strains previously mentioned was based on sequence similarity among the annotated proteins. Strains were chosen with specific characteristics to make an informed decision on the categorization of the isolated strains. PE-degrading strain E7, and LDPEdegrading strain PA01 as well as biosurfactant-producing strains UCBPP-PA14 and PA7 were chosen to understand how close the isolated strains are to these known plastic-degrading *P. aeruginosa* strains. Highly virulent PA14 strains were also chosen to find a distinction between plastic-degrading and virulent strains. Now, in this study, plastic degrading enzymatic pathways were searched from the WGS data to see if known enzymes and enzymatic pathways were available. For the polyethylene degrading enzymatic pathway, Yeom *et al.*, (2022) proposed a cytochrome P450 (P450)-driven cascade which starts with microbial deterioration of PE into alkane and is followed by hydroxylation by a hypothetical hydroxylase that may include P450. Then alcohol dehydrogenase, Baeyer-Villiger monooxygenase, and esterase work sequentially to produce alcohol and acid. From the sequences of this study, Cytochrome P450 107B1 gene, with P450-driven monooxygenase activity was found along with short-chain dehydrogenase, esterase, and Baeyer-Villiger monooxygenase respectively residing in the same operon. On the other hand, short-chain alcohol dehydrogenase and Baeyer-Villiger monooxygenase belonging to another operon were also found. Both of these operons seem suited for performing part of the degradation pathway. Additionally, alcohol dehydrogenases (*adh*) gene was also found which means most of the enzymes required in the pathway were present. Moreover, Cu-binding laccases, which are multi-copper oxidases, are predicted to be involved in the PE degradation pathway, as observed in previous studies (Fujisawa et al., 2001; Santo et al., 2013) and this enzyme was also present. Another enzyme, alkane hydroxylase or alkane monooxygenase (*alkB)* is also reported to be involved in PE degradation (Jeon & Kim, 2015) and genes for these enzymes were also found within the sequence.

When comparing these genes with other PE-degrading strains, orthologs were identified in most cases with over 99% sequence similarity. However, for PDB1_01395 and PDB1_01294, orthologs were not found in E7. Specifically, PDB1_01395 matched only with orthologs from PDB-2, PA01, and BWHPSA013, while PDB1_01294 matched only with orthologs from PDB-2 and PA01. This exclusivity may suggest a specific function for PE degradation performed by these enzymes, absent in other strains. The annotation of the PA7 strain identified a dyedecolorizing peroxidase (yfeX), corresponding to the protein PSPA7 2468 (GenBank Accession: ABR83505), previously labeled as a hypothetical protein. YfeX protein is associated with lignin peroxidases, which are predicted to be involved in the PE-degrading pathway (Mukherjee & Kundu, 2014). Unfortunately, there were no orthologous sequences of YfeX present in the isolated samples.

For EPS degradation, cleavage of the main chain can lead to styrene, which can then be broken down through two different metabolic pathways (Hou & Majumder, 2021). One of these requires styrene monooxygenase, and the other requires styrene dioxygenase. According to the Biocatalysis/Biodegradation Database (BBD) (Gao et al., 2010), in *Pseudomonas,* styrene monooxygenase, styrene oxide isomerase, and phenylacetaldehyde dehydrogenase work sequentially to convert styrene to phenylacetaldehyde, which is then passed on to the phenylacetate pathway. From the sequence, a short protein sequence annotated as phenylacetaldehyde dehydrogenase (*styD)* was found, but the coded protein was too short to be independently functional. For EPS degradation, it is known that polystyrene is a vinyl polymer and phenylacetaldehyde dehydrogenase is one of the most important enzymes to metabolize vinyl groups (Danso et al., 2019; Kiel et al., 2022). No genes were annotated as styrene monooxygenase and styrene oxide isomerase in this genome, nor could any be identified from the BLASTKoala search. P450 monooxygenases and alkane hydroxylases are also probable candidates to break the main-chain C-C bonds of PS, while the side-chain is perhaps cleaved by ring-hydroxylating dioxygenases (Hou & Majumder, 2021). According to previous studies, homoserine dehydrogenase (*hom*) and S-formylglutathione hydrolase (*yeiG*) genes also participate in plastic degradation, which is present in these isolates (H. R. Kim et al., 2020; H.-W. Kim et al., 2021). Other studies show that monooxygenases, alcohol dehydrogenases, and aldehyde dehydrogenases, all present in both strains, work together while degrading PE or PE-type plastics via an Oxo-degradation system (Gautam et al., 2007; Zeenat et al., 2021).

As previously mentioned there were several antibiotic-resistant genes were found, and they were in the bacterial chromosome. To prove they are expressed, AST was performed and showed the same result according to CLSI guidelines. Antibiotic resistance poses a problem while doing bioremediation, especially when *in situ*, as it increases the probability of transferring antibiotic resistance genes to other susceptible bacteria. However, plastic biodegradation can be more efficient with *ex-situ* methods, which will decrease the chance of spreading antibiotic resistance genes.

Virulence factors that are present in the isolates were analyzed and compared with other strains, and an important pathway involved in the production of rhamnolipid, a biosurfactant, was identified. Rhamnolipid is already in use for oil recovery, especially petroleum oils (Al-Sakkaf & Onaizi, 2023; Wei et al., 2020; Yan et al., 2012). Rhamnolipid biosynthesis has three genes, *rhlA, rhlB,* and *rhlC* (Toribio et al., 2010), and both isolates had all three genes. To prove these genes are expressed, biosurfactant assay was performed and all the systems had produced the rhamnolipid comparable to 10% Triton X-100. This result emphasizes the importance of these isolates in bioremediation, as they have potential utility for environmental recovery during oil spills.

As discussed earlier, PDB-1 isolate was identified from LDPE- and LLDPE-supplemented media, whereas PDB-2 isolate was sourced from EPS-supplemented media. However, it was not experimentally determined whether PDB-1 can grow on EPS and vice versa. MLST typing as well as comparison between protein sequences indicate that both isolates are

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phylogenetically not very distinct. In the future, further experiments may show that both strains are capable of degrading all three types of plastics.

Conclusion

Zophobus atratus larvae successfully survived on three different types of plastic, LDPE, LLDPE, and EPS, with varying rates of plastic degradation and survivability. FTIR analysis showed plastic oxidation occurring at the larval gut, indicating the gut microbiome was responsible. When bacteria isolated from the larval frass were cultured with plastics as the sole carbon source, growth was observed while SEM analysis (of the plastic samples incubated with isolated bacteria) confirmed the role of bacteria in plastics degradation. The isolated bacteria were later identified as *P. aeruginosa* and extensively characterized by whole genome sequencing. Through genome annotation, enzymes and enzymatic pathways involved in plastic degradation which have been observed in previous studies could be identified in the isolates. The presence of known virulence factors and antibiotic-resistance genes was also confirmed via sequence analysis. Further analysis is required to determine whether the candidate genes are expressed at a higher level when plastic is provided as a sole carbon source. Deliberate inactivation of these genes to create mutants will confirm their roles in the degradation process. Additionally, these isolates may be able to survive on other plastics, an aspect yet to be explored. In the current study, biological replication was performed in triplicate (n=20), and future research with a larger sample size should be performed to understand the biodegradation capacity of *Zophobus atratus* more extensively. Furthermore, the determination of the molecular weight of plastics (before and after biodegradation) is also recommended. Overall, the plastic-degrading bacteria in the larval gut provides an excellent opportunity to develop novel approaches for plastic bioremediation. From an environmental perspective, this study marks a starting point for a systems biology approach to design more potent enzymes and bacteria to combat plastic pollution.

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Appendix A

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